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2 **Seasonal variation in bee abundances in rangeland ecosystems of the Front Range region is**

3 **affected by land-use**

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10 **EXECUTIVE SUMMARY**

11 Researchers from CSU sampled pollinators using specialized trapping devices at three sites on
12 Boulder County Open Space and City of Boulder property in 2018. Results indicate significant
13 biodiversity of native bee pollinators in Boulder County, with bee abundances maximized in
14 June and July and in non-disturbed grassland habitats. Habitats exposed to cattle grazing
15 supported the fewest bees, but sites invaded with cheatgrass (*Bromus tectorum* or *B. japonicus*)
16 supported moderate bee abundances. Ongoing efforts on the project include bee identification
17 and development of species-area relationships, and continued updates will be provided to land
18 managers within the next calendar year.

19 **ABSTRACT**

20 Pollinator populations are experiencing widespread declines, and these losses are significant
21 concern for agriculturalists, food economies, and ecosystem management practitioners. Here, we
22 monitor native bee populations in Boulder County Open Space and City of Boulder sites in the
23 summer of 2018 using blue vane traps. A total of 368 specimens were collected over four
24 collection periods and represented by at least 17 taxa with ~38% of the overall bee abundance
25 represented by bumblebees (*Bombus* spp.) and ~20% of the overall abundance comprised of
26 longhorn bees (*Melissodes* spp.). Bee population abundance and species richness was compared
27 across seasons, with peak abundance and diversity in June, though July also supported high bee
28 abundances. Sites were in grassland habitats and consisted of a grazed site, a cheatgrass-invaded
29 site, and a relatively non-disturbed 'natural' grassland site. Abundances were lowest overall in
30 the grazed site, highest at the non-disturbed site, and intermediate at the cheatgrass-invaded site;
31 these differences were particularly pronounced early in the growing season (May-Jul). However,
32 late in the growing season (Aug), bee abundances at grazed sites resembled those at non-
33 disturbed and cheatgrass-invaded sites. This monitoring effort is part of a larger sampling
34 network of 30 total sites spanning the Front Range region; ongoing efforts are focusing on
35 continuing to characterize taxonomic diversity in bee specimens across all sites and make
36 comparisons of seasonal variation in bee abundance and diversity across grazed, invaded, and
37 non-disturbed habitats. Submission of a larger and more detailed report as a peer-reviewed article
38 is expected in 2019.

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40 **Keywords:** Biodiversity, Grassland habitats, Land Use Classification, Pollinators, Rangeland

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INTRODUCTION

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Wild pollinators are key species for agriculture as well as natural vegetation playing vital role in maintaining ecosystem biodiversity and function (Garibaldi et al., 2013, Kremen et al., 2007). Approximately 75% of crop-plant species globally depend on pollinators (Kleijn et al., 2015, Ollerton et al., 2011), and despite their crucial role in ecosystem productivity, there remains little known about factors regulating abundance or diversity of insect pollinators in Colorado or on rangelands in general (Gilbert and Vaughan 2011), though pollinator fauna are described for some urban landscapes (Kearns and Oliveras 2009a, b; Scott et al. 2011).

Our study sites consist primarily of shortgrass prairie rangelands along the foothills of Front Range of Colorado. The shortgrass prairie is dominated by blue gramma (*Bouteloua gracilis*), buffalo grass (*Bouteloua dactyloides*) and others. Besides grasses, rangelands harbor many native forb species crucial to pollinators including *Liatris punctuata*, *Grindelia squarrosa*, prickly pear cacti (*Opuntia polyacantha*), and rabbitbrush (*Ericameria nauseosa*), among others. It is considered that vegetation in the rangelands are evolved with grazing and have developed tolerance to ungulate grazing (Klippel and Costello, 1960) and many rangelands are naturally grazed by bison, antelope, and other species. Now, most of Front Range rangelands area either privately and publicly ranches or open space natural areas developed for recreational activities. Most open space areas are enclosed or otherwise protected from commercial grazing by park or county resource managers. Though most of the open spaces are enclosed from grazing, many are invaded by cheatgrasses (invasive *Bromus* spp.) native to Europe and Asia. Evidence has shown that these grasses are expanding in rangelands of the western United States and altering both vegetation communities and natural disturbance regimes (Knapp1996). Accordingly, the

65 Colorado Front Range is an ideal landscape to study the effects of both grazing and invasive
66 species (cheat-grass) on interactions between habitat factors and pollinator communities.

67 It is presently unknown whether invasion by cheatgrasses or grazing are associated with
68 wild pollinator declines in the Front Range region. Exotic plant species have indirect effect on
69 pollinators by changing the distribution, composition and abundance of native floral resources
70 and/or nesting habitats (Goodell and Parker 2017). For instance, site invasion by *Bromus*
71 *tectorum* is widespread in Front and may reduce local site occupancy by forbs (Parkinson et al
72 2013) or constrain local species richness (Melgoza et al, 1990, Knapp 1996); these effects may
73 indirectly effect bee communities by altering access to foraging and nesting sites (McKinney and
74 Lockwood, 1999). Here, our goal is to compare the relative effects of cheatgrass invasion and
75 cattle grazing on bee species abundance and richness, as well as the abundance and richness of
76 floral resources, using relatively nondisturbed grassland habitats as ‘control’ sites. Sites in
77 Boulder County represent approximately 10% of our sampling effort across the Front Range, and
78 our short-term results provide new insights into seasonal variation in pollinator abundance across
79 three sites.

80 METHODS

81 **Site selection and field measurements:** We selected three habitats types for sampling;
82 sites that were (1) grazed by cattle, (2) sites that were dominated by cheatgrass, and (3) relatively
83 non-disturbed ‘natural’ grassland sites (Table 1). All sites are a minimum distance of 1 km to
84 reduce potential for spatial autocorrelation, and one site of each ‘type’ is located in Boulder
85 County. A geographic information system (ARCGIS, Esri, Inc.) was used to characterize
86 landscape diversity based on remote sensing data (LANDFIRE datasets), and the relative
87 proportion of each site that was ‘urban’ within a 500 m buffer zone around each site was

88 quantified. At each site, point-intercept transects were used to characterize site vegetation, with
89 five 100-m transects placed equidistant around bee traps, and intercepts placed at 1 m distances
90 along each transect. At each intercept, cover was determined to be ‘native grass’, ‘*Bromus* spp.’,
91 ‘floral’, ‘woody debris’, ‘rock’ or ‘soil’. All floral intercepts were identified to the highest level
92 of taxonomic resolution (species, in many cases), and floral traits were quantified for each
93 identified species.

94 **Bee collection and identification:** Bees were collected using blue vane traps (Springstar,
95 Inc., Woodinville, WA). Collections were made by placing one trap at a central location in study
96 sites, with the trap base suspended ~1 m from the ground surface. Traps were left for a period of 72
97 h and then contents were collected into large plastic bags, placed on ice immediately, and
98 returned to the laboratory on the same day. One collection was made at each site in May, June,
99 July, and August in 2018 (Table 2). Trap contents were sorted, and all bees were pinned, labeled,
100 and mounted in collection boxes. Bees were sorted to morphospecies and identified to the
101 highest level of taxonomic resolution by comparing to specimens curated in the C.P. Gillette
102 Museum (CSU), as well as using pollinator identification resources (textbooks) and the expertise
103 of professional insect taxonomists (Prof. Boris Kondratieff, CSU). Identification is ongoing and
104 projected for completion in Spring of 2019.

105 RESULTS

106 A total of 368 specimens were collected over four collection periods and represented by
107 at least 17 taxa with ~38% of the overall bee abundance represented by bumblebees (*Bombus*
108 spp.), ~20% of the overall abundance comprised of longhorn bees (*Melissodes* spp.), and ~14%
109 of bee abundance comprised of sweat bees (*Lasioglossum* spp.; Figure 1; more detailed species
110 list provided in Table 3). Bee population abundances were highest in June and July, with abrupt

111 declines in August, likely due to initiation of overwintering behaviors. Through the early- (May)
112 and middle-growing season (June, July), the site with cattle grazing exhibited the lowest overall
113 bee abundance, but this difference was eroded by late in the growing season (August). Non-
114 disturbed control grassland sites supported the greatest bee abundance, and sites invaded by
115 cheatgrasses were intermediate (Figure 2)

116 **DISCUSSION**

117 In our preliminary monitoring of bee communities in Boulder County, we determined
118 that significant bee diversity exists within the county, and that this diversity tends to vary
119 seasonally and relative to the type of biotic disturbance that occurs at sample sites (cattle grazing
120 vs. cheatgrass invasion vs. non-disturbed grassland). Altogether, a relatively small proportion
121 (2%) of the bees collected in this study It remains undetermined whether these findings are
122 broadly extendable to sites across the region, and a clearer understanding of these biotic effects
123 on pollinator resources is pending final identification of specimens and a more thorough analysis
124 of bee community data. More detailed findings are projected by Spring 2019, and final reporting
125 will take the form of a peer-reviewed manuscript to be submitted for publication.

126 It is not entirely unexpected that grazing resulted in reduced bee abundances, as
127 consumption of forb species by cattle may be a primary mechanism for loss of biodiversity in
128 rangeland systems (). However, it was surprising that cheatgrass-invaded sites maintained a
129 higher species abundance than grazed sites, as these sites are generally depauperate of floral
130 resources early in the growing season. It is possible that cheatgrass-invaded sites are superior
131 nesting sites, but not foraging sites. Another possibility is that landscape richness and
132 urbanization, with are both processes occurring at larger spatial scales than at the sites of bee
133 collection, are driving bee abundances and diversity more so than effects occurring at the site-

134 scale. Further analysis will reveal whether local or landscape factors are stronger drivers of local
135 abundances and diversity in Front Range bee communities.

136 Understanding the factors that govern local bee abundances and diversity will be
137 fundamental to designing conservation strategies for retention of ecosystem services in the Front
138 Range, particularly with continued population growth and projections of ongoing urbanization.
139 Our results, once finalized, can help managers to determine which habitat types may be more
140 likely or less likely to support pollination services, and will provide specific locations (i.e.,
141 sample collection sites) where directed vegetation management actions may be useful for
142 enhancing pollinator diversity. In addition, future work will characterize functional diversity of
143 bee communities and floral resources at study sites and is expected to provide an additional
144 dimension of analysis. For instance, functional diversity may be equivalent across habitat types,
145 even if biological or taxonomic diversity is variable across habitat types.

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169 Apoidea: Anthophila). Natural History Inventory of Colorado. University of Colorado
170 Museum of Natural History, Boulder, Colorado. 112 p.

171 **Table1:** Study sites with their coordinates.

Sites	Coordinates (°N, °W)
Agriculture center	40.201236, 105.162970
Ron Steward Preserve at Rabbit Mountain (South)	40.250538, 105.214050
Ron Steward Preserve at Rabbit Mountain	40.241479, 105.225994

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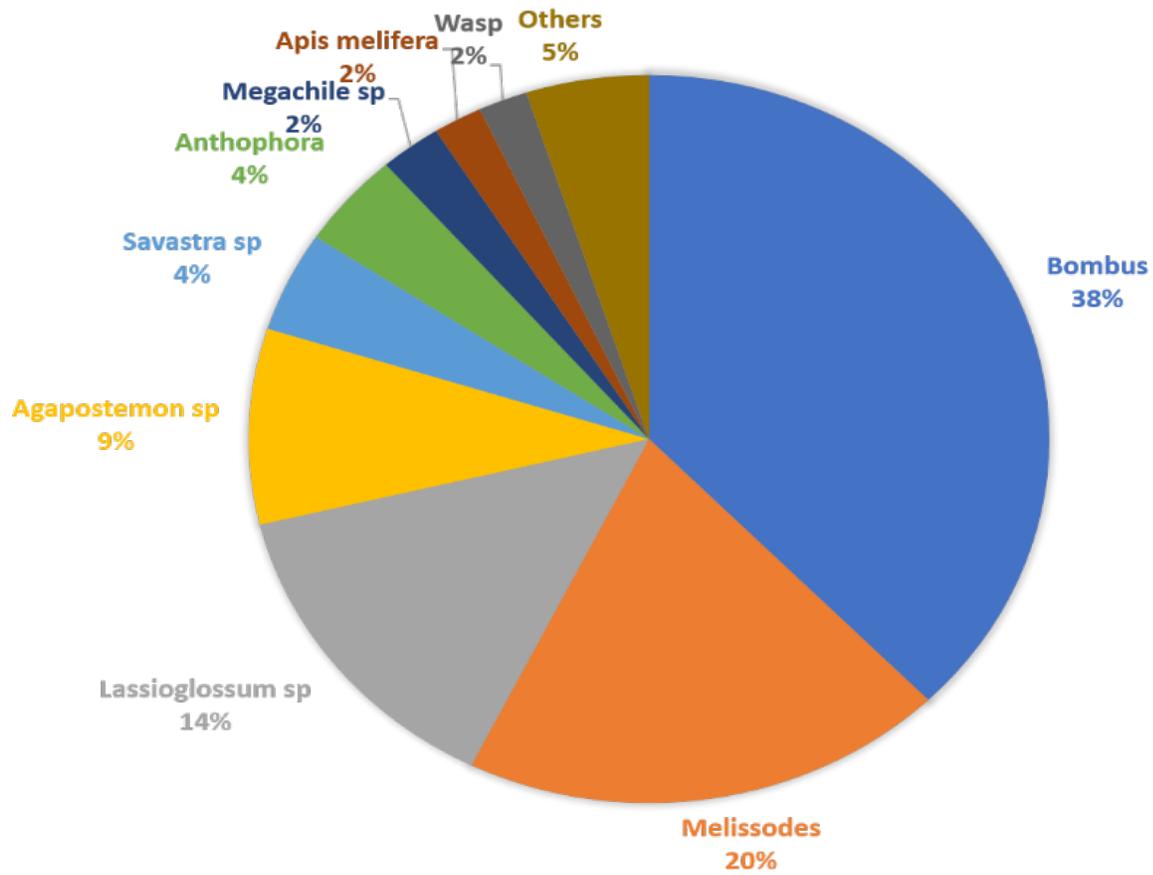
173 **Table 2.** Installation and collection dates for blue vane traps in 2018.

BVT installation date	BVT collection date
May-16, 2018	May-19, 2018
June-18, 2018	June-22, 2018
July-18, 2018	July-21, 2018
August-10, 2018	August-13, 2018

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175 **Table 3.** Abundances of identified bee taxa from field collections in 2018. Current taxonomic
 176 resolution is coarse, and diversity is expected to increase substantially as sorted morphospecies
 177 are identified by expert entomologists.

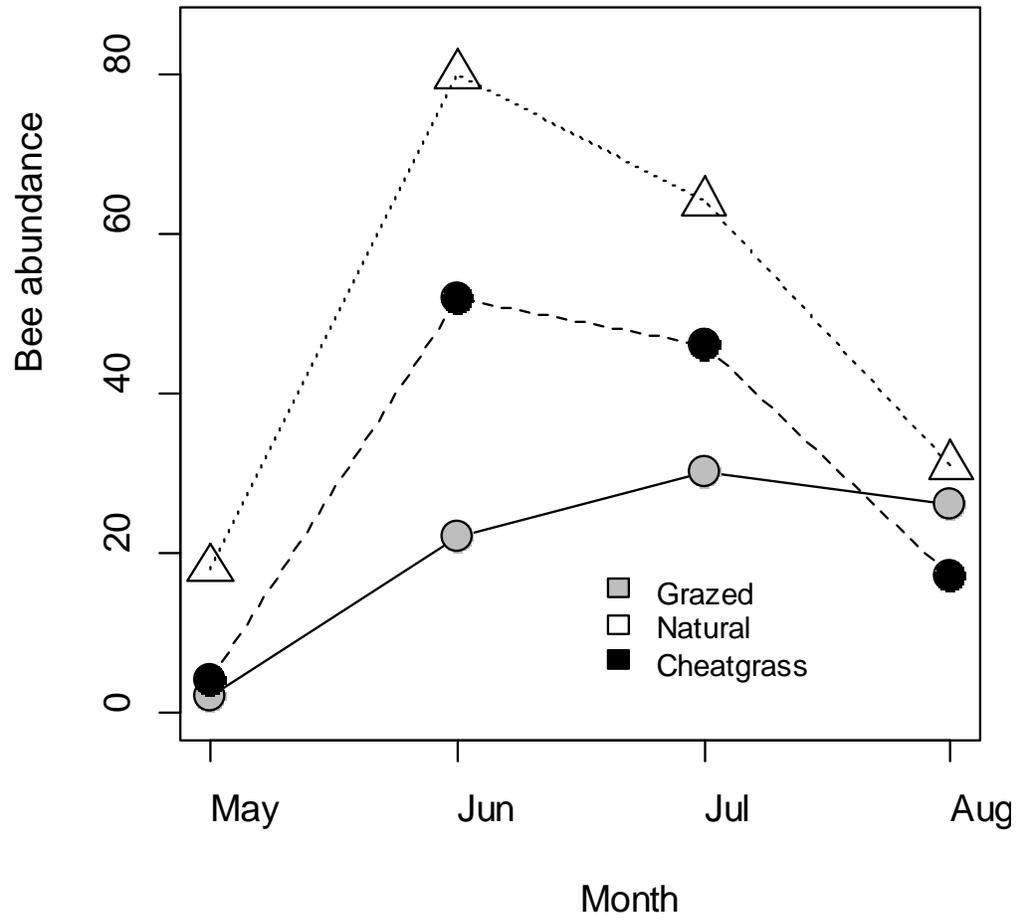
Species	No. individuals
<i>Agapostemon</i> sp	35
<i>Anthidium</i> sp	4
<i>Anthophora</i> sp	17
<i>Apis mellifera</i>	8
<i>Bombus appositus</i>	7
<i>Bombus fervidus</i>	28
<i>Bombus griseocollis</i>	5
<i>Bombus huntii</i>	2
<i>Bombus nevadensis</i>	21
<i>Bombus pensylvanicus</i>	87
<i>Lasioglossum</i> sp	56
<i>Megachile</i> sp	10
<i>Melecta pacifica</i>	2
<i>Melissodes</i> sp	79
Other species/not identified	16
<i>Svastra</i> sp	18
<i>Vespula pensylvanica</i>	8
SUM	<u>368</u>



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181 **Fig. 1.** Relative composition of bee genera trapped during the 2018 field season on Boulder

182 County Open Space and City of Boulder study sites.



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184 **Fig. 2.** Seasonal variation in bee abundances at grazed, cheatgrass-invaded, or 'natural', sites.